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Biogenic habitat structure of seaweeds change along a latitudinal gradient in ocean temperature

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ABSTRACT

Global warming is affecting all major ecosystems, including temperate reefs where canopy-forming seaweeds provide biogenic habitat. In contrast to the rapidly growing recognition of how climate affects the performance and distribution of individuals and populations, relatively little is known about possible links between climate and biogenic habitat structure. We examined the relationship between several ocean temperature characteristics, expressed on time-scales of days, months and years, and habitat patch characteristics on 24 subtidal temperate reefs along a latitudinal gradient (Western Australia; ~34 to 27°S). Significant temperature-related variation in habitat structure was observed, even though the total landscape cover of kelp and fuclean canopies did not change across the climate gradient: monospecific patches of kelp became increasingly dominant in warmer climates, at the expense of mixed kelp–fuclean canopies. The decline in mixed canopies was associated with an increase in the abundance of *Sargassum* spp., replacing a more diverse canopy of *Scytothalia doryocarpa* and several other large furoids. There were no observed differences in the proportion of open gaps or gap characteristics. These habitat changes were most closely related to patterns in minimum temperatures and hot temperature thresholds (days >20 °C), presumably because temperate algae require cool periods for successful reproduction and recruitment (even if the adults can survive warmer temperatures). Although the observed habitat variation may appear subtle, similar structural differences have been linked to a range of effects on canopy-associated organisms through the provision of habitat and ecosystem engineering. Consequently, our study suggests that the magnitude of projected warming might cause changes in habitat structure and thereby indirectly affect numerous habitat-dependent plants and animals.

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1. Introduction

Global warming has already affected all major ecosystems on earth (Rosenzweig et al., 2008) including temperate nearshore habitats where temperature increases of >0.5–1 °C have been measured over the past decades (Hawkins et al., 2003; Pearce and Feng, 2007; Ridgway, 2007). Whilst the current understanding and documentation of the physical changes that are taking place are fairly good, the understanding of how this physical forcing translates into effects in the biological world is comparatively poor, particularly in the sea (Harley et al., 2006; Poloczanska et al., 2007). Yet, this information is critical to understanding the range of consequences of global warming, and how they might be mitigated.

Foundation species are critical to determining overall community structure in the habitats where they are found (Dayton, 1972). Canopy-forming macroalgae are the main foundation species on most temperate rocky reefs, where two groups, kelps (order Laminariales) and furoids (order Fucales), are particularly abundant (Schiel and Foster, 1986). The mechanisms behind their influence on community structure are multifaceted, but key ecological functions include modifying local environmental conditions (e.g., Jenkins et al., 1999a,b,c; Wernberg et al., 2005; Irving and Connell, 2006b), and the provision of habitat, including a 3-dimensional physical structure that can be inhabited by various organisms (e.g., Thompson et al., 1996; Wernberg et al., 2004; Tuya et al., 2008; Thomsen et al., 2010). Elevated temperature negatively affects most aspects of the biology of kelps and furoids from temperate coasts (e.g., reproduction, recruitment and growth, and resilience to disturbances, Kirkman, 1984; Novaczek, 1984; Wernberg et al., 2010). Thus, projected global warming is causing serious concern for permanent changes in habitat structure and loss of algal habitats (Kendall et al., 2004; Poloczanska et al., 2007; Hawkins et al., 2008, 2009) with indirect effects on other community components (Schiel

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et al., 2004) and ensuing loss of associated biodiversity (Graham, 2004; Ling, 2008; Blight and Thompson, 2008) and cascading effects to higher trophic levels (Kendall et al., 2004; Moore et al., 2007; Ling et al., 2010). Indeed, habitat loss compounded by climate change, is probably the biggest threat to biodiversity and ecosystem function, in general (Pimm, 2008).

Despite concerns for the future persistence of algal habitats and their ecological functions, there have been few direct assessments of how algal habitats (as opposed to species or individuals) might respond to increasing water temperatures (Hawkins et al., 2009; Merzouk and Johnson, 2011; Schiel, 2011-this issue). Broad-scale oceanographic events causing elevated ocean temperatures (i.e., ENSO's on the west coast of the America's) have consistently been associated with loss and fragmentation of kelp canopies and substantial changes to biogenic habitat structure (Dayton et al., 1999; Martinez et al., 2003; Edwards and Estes, 2006). The influence of temperature is, however, often ambiguous because of a strong negative relationship with water-column nutrient concentrations in many kelp-dominated systems (particularly in areas of upwelling, e.g., Edwards and Estes, 2006). In a unique case study of the bay-wide effects of 10 years of induced heating by thermal effluent from a power plant, Schiel et al. (2004) found substantial structural changes associated with a temperature increase of 3.5 °C; they reported large shifts in habitat formers where kelps, for example, decreased markedly in abundance. With the exception of such unique 'opportunistic experiments', it is impossible to manipulate temperature in subtidal habitats across spatial scales relevant to questions of landscape structure (1–10 m, Connell and Irving, 2008). Questions about the influence of water temperature on habitat structure must, therefore, follow alternative lines of enquiry. One option is to study latitudinal gradients in ocean temperature at places where confounding by other co-varying factors is minimal (e.g., Wernberg et al., 2010). Understanding how latitudinal gradients in temperature influence the distribution of key organisms is particularly relevant in relation to predicting effects of climate change, because shifts in distribution patterns along these gradients have been one of the main detectable responses of both aquatic and terrestrial organisms (Southward et al., 1995; Walther et al., 2002; Parmesan and Yohe,

2003; Lima et al., 2007; Poloczanska et al., 2007; Hawkins et al., 2008, 2009; Wernberg et al., 2011).

Here, we took advantage of a latitudinal gradient in ocean climate to explore the influence of temperature on biogenic habitat structure, defined as the relative abundance of key algal habitats and their patch size characteristics. In doing so, we tested the hypothesis that habitat structure would change gradually from latitude to latitude, and that the dominant kelp habitats would become increasingly fragmented at lower latitudes where ocean temperatures are higher, and presumably more stressful to temperate algae.

2. Methods

2.1. Study area and experimental design

Our study took place along a ~1000 km stretch of the southwest coast of Western Australia, covering the latitudes 34°S to 27°S (Fig. 1A). This coastline runs north–south and its near-shore environment is characterised by limestone and sandstone reefs. The ocean climate is dictated by the Leeuwin Current, a warm surface current which, in contrast to other eastern boundary currents in the southern hemisphere, runs southward and therefore suppresses upwelling (Pearce, 1991). The result is a latitudinal temperature gradient of 2–3 °C where nutrient concentrations remain consistently low (Pearce, 1991; Lourey et al., 2006). This temperature gradient captures the range of current scenarios for global warming of Australia's temperate waters (25-year projection of 1–2 °C increase and 50-year projection of 2–4 °C, Poloczanska et al., 2007). Grazing pressure from fish and invertebrates is also low throughout this region (Vanderklift and Kendrick, 2005; Wernberg et al., 2008; Vanderklift et al., 2009).

The reefs along the coastline host a diverse algal flora dominated by a canopy of the only kelp in Western Australia, *Ecklonia radiata* (a species similar to northern kelps such as *Eisenia arborea* and *Laminaria digitata*) and a rich assemblage of fuclean algae mainly from the genera *Cystophora*, *Sargassum* and *Scytothalia* (Wernberg et al., 2003; Smale et al., 2010). Similar reef habitats and canopies characterise

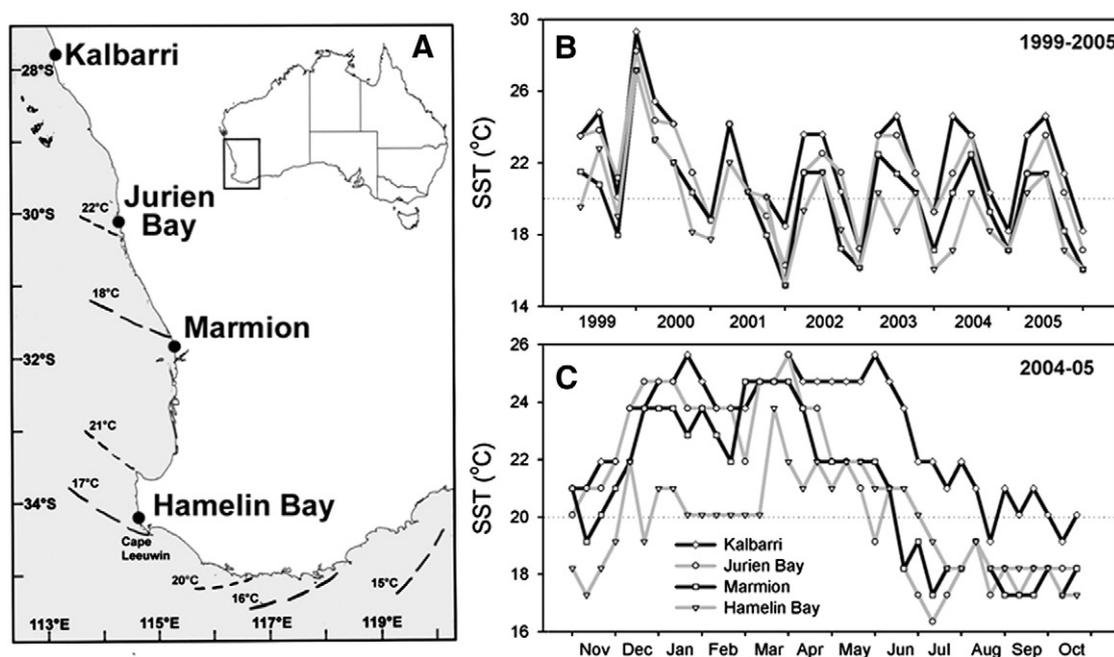


Fig. 1. A: Position of the four regions included in this study, on the southwest coast of Western Australia. Summer (short dash) and winter (long dash) isotherms after Pearce (1991). B and C: Sea Surface Temperature (SST) measured by remote sensing (AVHRR). The dotted lines indicate 20 °C, a threshold for when the growth and productivity of *Ecklonia radiata* start to decline (Kirkman, 1984; Hatcher et al., 1987).

rocky reefs throughout temperate Australasia (Wernberg et al., 2003; Connell and Irving, 2008).

All sampling was completed within one month (November 2005) and followed a hierarchical design with four regions (each representing a different temperature regime) evenly separated by $\sim 2^\circ$ latitude (Fig. 1A). Six independent reefs, 8–12 m deep and separated by a minimum of 1 km, were nested within each region. The depth (transect start and end) and structural complexity (count of intersecting topographic features of 1–2 m and >2 m vertical relief) were measured for 25 transects (see below) for each reef. There were no differences in depth or structural complexity amongst regions ($MS_{depth} = 5.61$, $F_{(3,20)} = 1.0$, $P = 0.428$; $MS_{LN(1-2m+1)} = 0.56$, $F_{(3,20)} = 0.96$, $P = 0.430$; $MS_{LN(>2m+1)} = 0.24$, $F_{(3,20)} = 0.58$, $P = 0.634$). Also, differences in light amongst regions due to latitude (5–10%) were masked by differences in light ($\sim 25\%$) due to depth differences amongst reefs within regions (see calculations in Stæhr and Wernberg, 2009).

Here, each reef was considered a biologically independent replicate of the temperature regime represented by each region. This was justified because subtidal temperatures vary little ($<5\%$) between reefs at similar depth (Smale and Wernberg, 2009), and because most propagules from canopy-forming macroalgae have short dispersal distances (metres or less, e.g. *Sargassum spinuligerum*: Kendrick and Walker, 1991) resulting in low genetic connectivity between reefs (e.g. *E. radiata*: Coleman et al., 2009).

2.2. Ocean temperature

The unique ocean climate, and the broad latitudinal patterns of temperature change along the southwest coast, have long been established (Pearce, 1991, Fig. 1A). Recently, Smale and Wernberg (2009) showed strong correlations between temperatures at 8–10 m depth and satellite-derived sea surface temperatures (SST) in this area. Consequently, the temperature regime of each region at different times prior to measurement of habitat structure could here be inferred from historic satellite derived SST's.

SST's obtained by the 'Advanced Very High Resolution Radiometer' (AVHRR) aboard the 'National Oceanic and Atmospheric Administration' (NOAA) series of satellites have been archived by the Australian Bureau of Meteorology. Daily SST maps of Australian waters were used to generate temperature profiles for each of the four study regions (Australian Bureau of Meteorology, www.bom.gov.au) (see Smale and Wernberg, 2009 for details). Two data sets were generated: one covering seven years (1999–2005, assessed every 15th of January, April, July and October) and one covering 1 year (November 2004–October 2005, assessed every 1st, 10th and 20th of each month). Ambient water temperature was also measured at the time of sampling (Onset Tidbit pendant loggers, logging every 5 min ~ 15 cm above the bottom at all six reefs within each region). The combination of these measures integrates long-term temperature exposure across >1 full generation of *E. radiata* (life span 1–4 years, Wernberg, 2005), and more short-term effects due to the particular year, season and time of sampling. The mean temperature, maximum and minimum temperatures, and number of 'hot' days $>20^\circ\text{C}$, were used to characterise chronic thermal exposure, extreme thermal exposure, and 'adverse' thermal exposure, respectively; 20°C corresponds to the approximate threshold where temperature starts to negatively affect the productivity and growth of *E. radiata* (Kirkman, 1984; Hatcher et al., 1987).

2.3. Habitat structure

The overall biogenic habitat structure of each reef was quantified by running ten 25 m transects along haphazard compass bearings and recording the distances (to nearest 10 cm) of habitat change. This technique has previously been successfully applied to describe habitat

structure in algal dominated assemblages in other subtidal habitats (Kennelly, 1987; Wernberg, 2006; Connell and Irving, 2008). We distinguished between four major habitat types: (1) mono-specific kelp canopy ($>90\%$ *E. radiata*), (2) fuclean canopy ($>50\%$ fuclean algae), (3) mixed kelp and fuclean canopy (50–90% kelp and 10–50% fuclean algae) and (4) canopy-free areas (gaps of $<10\%$ canopy algae), typically dominated by small foliose red algae (<25 cm) (e.g., *Hennedya crispa*) or articulated coralline algae (e.g., *Amphiroa anceps*). The distinction between these habitat types was based on previous studies documenting how they affect their immediate surroundings differently, and support different associated benthic, phytal and fish communities (Harman et al., 2003; Irving et al., 2004; Wernberg et al., 2005; Irving and Connell, 2006a; Tuya et al., 2008). Habitat structure was sampled in November 2005 (early Austral summer), a time where the kelp canopy has recovered from the thinning and pruning of winter storms (Wernberg and Goldberg, 2008).

Overall, 16 descriptors of habitat structure were considered, four for each habitat type: habitat cover was calculated as the proportion of each transect occupied by each habitat type. Median and maximum habitat patch sizes (1-dimensional linear distance) as well as the number of habitat patches for each habitat type were determined for each reef after pooling all transects.

2.4. Statistical analyses

Differences in mean ocean temperature characteristics between adjacent regions were tested by either ordinary or paired *t*-tests (Zar, 1996). Overall patterns in ocean temperature were assessed by principal components analysis (PCA, Clarke and Gorley, 2006) based on all nine temperature characteristics (cf. Table 1), and the presence of sequential change from region to region tested by non-parametric serial correlation (RELATE, Clarke and Gorley, 2006).

Differences in biogenic habitat structure amongst regions, taking all 16 habitat characteristics into consideration, were tested by multivariate analysis of variance by permutation (PERMANOVA, Anderson et al., 2008). Constrained ordination by canonical analysis of principal coordinates (CAP, Anderson et al., 2008) was used to characterise, and visualise, these differences, and to evaluate their correspondence with individual temperature variables in the environmental matrix by canonical correlation (CCorA, Anderson et al., 2008). These analyses were based on normalised data for habitat characteristics, using Euclidian distances and 999 permutations.

Subsequently, the influence of ocean temperature on biogenic habitat was tested by linear regression (Zar, 1996) of each habitat characteristic against PC1, the first principal components axis from the PCA of patterns in ocean temperature. PC1 represents a univariate proxy for multivariate variation across all ocean temperature characteristics; this approach recognises the potential concurrent influence of multiple temperature characteristics as drivers of patterns in different habitat elements. Regression analyses were carried out for all four habitat characteristics (maximum patch size not shown, as $P > 0.1$ in all cases, cf. Fig. 3) of all four habitat types, and for total cover of four dominant canopy taxa.

3. Results

3.1. Ocean temperature

There were clear differences in ocean temperature regimes along the coastline, and the overall pattern for almost all temperature characteristics was an increase from Hamelin Bay in the south towards Kalbarri in the north (Fig. 1, Table 1). Ocean temperatures showed a clear seasonal pattern of ~ 6 – 8°C variation between summer and winter in all four regions. Inter-annual differences were generally small, particularly after 2000 (Fig. 1B). The four regions spanned a temperature gradient of ~ 2 – 3°C ; Hamelin Bay was

Table 1

Ocean temperature characteristics of the regions studied. Temperature data for time of sampling were measured *in situ* where habitat characteristics were sampled. Temperature data for 2004–2005 and 1999–2005 were obtained by remote sensing (see [Methods](#)).

Temperature characteristics	Hamelin Bay	Marmion	Jurien Bay	Kalbarri	Pattern
<i>Time of sampling (Tidbit data logger)</i>					
Mean (°C, ± SE)	17.95 ± 0.03	19.34 ± 0.11	19.97 ± 0.13	21.21 ± 0.11	H < M < J < K ($P < 0.001$) ^a
<i>2004–05 (SST, AVHRR)</i>					
Mean (°C, ± SE)	19.82 ± 0.27	20.93 ± 0.43	21.08 ± 0.48	22.94 ± 0.34	H < M = J < K ($P < 0.001$) ^b
Minimum (°C)	17.28	17.28	16.35	19.14	H = M = J < K
Maximum (°C)	23.79	24.72	25.65	25.65	H < M < J = K
Hot days > 20 °C (%)	55.6	61.1	61.1	94.4	H < M = J < K
<i>1999–2005 (SST, AVHRR)</i>					
Mean (°C, ± SE)	19.48 ± 0.50	20.13 ± 0.51	21.42 ± 0.52	22.10 ± 0.53	H < M < J < K ($P < 0.010$) ^b
Minimum (°C)	15.18	15.18	16.28	17.22	H = M < J < K
Maximum (°C)	27.16	27.16	28.24	29.32	H = M < J < K
Days > 20 °C (%)	46.4	60.7	71.4	78.6	H < M < J < K

^a Ordinary *t*-tests.

^b Paired *t*-tests, paired by sampling dates.

substantially cooler in summer (Dec–Feb) whereas Kalbarri remained substantially warmer in winter (Jun–Oct) (Fig. 1C). Conditions at Marmion and in Jurien Bay were intermediate (Table 1).

Considering all temperature characteristics simultaneously, PCA revealed a strong serial correlation in temperature regimes from region to region (RELATE, $\rho = 0.926$, $P = 0.081$, $n = 4$), emphasising the gradual environmental change along the southwest Australian coastline. The first principal component (PC1) accounted for 85.2% of the multivariate differences in ocean temperatures, and thus provided a strong univariate proxy for overall variation in temperature regime amongst regions, where larger values were associated with warmer (more stressful) conditions.

3.2. Habitat structure

Based on all 16 patch descriptors, there were significant differences in biogenic habitat structure between regions (PERMANOVA, pseudo- $F = 2.36$, $P = 0.020$). The constrained ordination grouped all four regions into distinct groups (CAP, *trace statistic* = 0.410, $P = 0.030$, Fig. 2) where Kalbarri, Marmion and Jurien Bay were separated along CAP1 and Hamelin Bay from these three groups along CAP2. Canonical correlation (CCorA) against the temperature matrix revealed a significant relationship between temperature regime and biogenic habitat structure (CAP, *trace statistic* = 1.062, $P = 0.003$), where minimum temperature and the number of hot days in the year preceding sampling (2004–2005) were the best correlates of CAP1

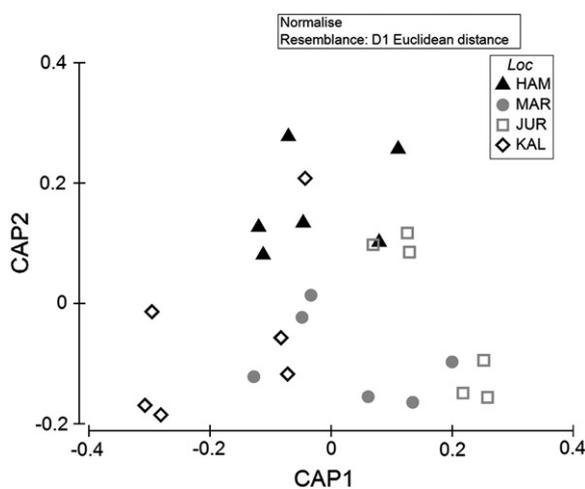


Fig. 2. Constrained ordination of principal coordinates for habitat structure at six reefs within Hamelin Bay (HAM), Marmion (MAR), Jurien Bay (JUR) and Kalbarri (KAL).

and temperature at the time of sampling, maximum temperature in the preceding year, and the number of hot days during the preceding 7 years (1999–2005), were the best correlates of CAP2 (Table 2). These temperature characteristics have, therefore, likely been influential in generating the observed patterns of biogenic habitat structure.

All four habitat types were present in all four regions, although only monospecific kelp canopies and open gaps were present at all 24 reefs (Fig. 3). Mono-specific kelp canopies covered 20–90% of the reefs at all sites, and were by far the dominant habitat type. In contrast, fuclean canopies covered <30% of the reefs and were generally the least common habitat type. With a median size of 2–5 m, kelp patches were also the largest, followed by mixed canopy patches (~2–3 m), fucoid canopies (~1–3 m) and open gaps (~1–1.5 m).

There was substantial variation in most biogenic habitat characteristics both amongst and within ocean temperature regimes (as represented by regions) (Fig. 3). For example, the cover of monospecific kelp canopies ranged from ~10 to 70% on reefs within Hamelin Bay and mixed kelp and fucoid canopies from 0 to 30% in Kalbarri. Despite this variation, there was a significant positive relationship between PC1 for ocean temperature and monospecific kelp canopy cover and median size of kelp patches, where ocean temperature explained 21% and 34% of variation, respectively (Fig. 3). There was a near-significant trend for a negative relationship between ocean temperature and cover and patch density of mixed canopies, with temperature accounting for 13% and 15% of variation respectively ($P < 0.081$). Ocean temperature explained <6% of variation in habitat characteristics related to fucoid canopies, open gaps and maximum patch size (not shown).

Despite a positive relationship between ocean temperature and cover of monospecific kelp canopies, the total cover of *E. radiata* (i.e. including kelp found within mixed canopies) did not correlate with

Table 2

Canonical eigenvectors (correlation coefficients) for individual ocean temperature characteristics in relation to the constrained ordination of biogenic habitat structure on reefs within regions. The five best correlated variables are highlighted in bold.

Temperature variable	CAP1 ($\delta_1^2 = 0.54$)	CAP2 ($\delta_2^2 = 0.45$)
Temperature, time of sampling	-0.152	- 0.486
Mean temperature 2004–2005	-0.305	-0.399
Minimum temperature 2004–2005	- 0.636	0.060
Maximum temperature 2004–2005	0.083	- 0.474
Percent hot days > 20 °C, 2004–2005	- 0.489	-0.038
Mean temperature 1999–2005	-0.139	-0.107
Minimum temperature 1999–2005	-0.314	0.273
Maximum temperature 1999–2005	-0.336	0.268
Percent hot days > 20 °C, 1999–2005	-0.052	- 0.465

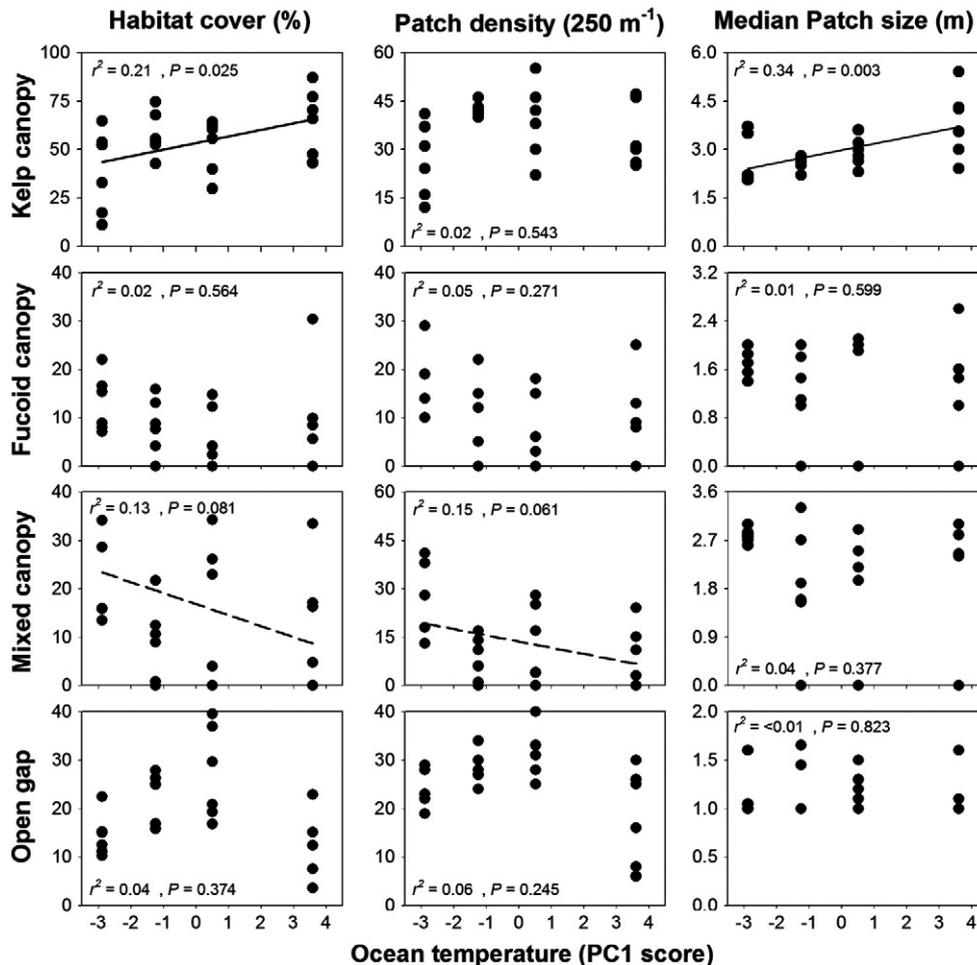


Fig. 3. Relationship between ocean temperature at four regions (from left to right: Hamelin Bay, Marmion, Jurien Bay and Kalbarri) and characteristics of the biogenic habitat at 6 independent sites within each region. Kelp canopy refers to mono-specific patches of *E. radiata*. The ocean temperature incorporates temperature variation at multiple temporal scales (cf. Table 1). Solid lines indicate significant relationships ($P < 0.05$) and dashed lines near-significant relationships ($0.05 < P < 0.10$).

ocean climate (Fig. 4, ocean temperature explained <5% of variation in total kelp cover). However, where temperature explained little of the variation in the combined fucooid abundance (Fig. 3), the different fucooid taxa still showed significant, but opposing, relationships with temperature, explaining 21–37% of the variation of these taxa (Fig. 4). In particular, where the relationship was positive for *Sargassum* spp., it was negative for *Scytothalia doryocarpa* canopies and for ‘other fucooids’.

4. Discussion

Our study has shown a significant relationship between the biogenic habitat structure of temperate subtidal algal canopies and ocean temperature along a latitudinal gradient. The total cover of kelp and fucooids did not change along the gradient, but changes to canopy structure and individual fucalean taxa were evident. In contrast to expectation, kelp canopies did not become increasingly fragmented in warmer, more stressful, temperature regimes. Rather, mono-specific kelp canopies became increasingly dominant as a result of a (narrowly non-significant) decline in mixed kelp–fucalean canopies, and an opposing relationship to temperature of individual fucalean taxa. Attributing these changes solely to variation in ocean temperature is, of course, impossible because of the correlative nature of this study and the potential for covariation with other important drivers along the latitudinal gradient. However, in our study area, there is a particularly low degree of confounding between the temperature gradient and other environmental factors known to drive ecological patterns across latitudinal gradients

(discussed in Smale and Wernberg, 2009; Wernberg et al., 2010), and we therefore believe ocean temperature has a pivotal role as a driver of the observed patterns.

4.1. Ocean temperature

The unique oceanography of the southwest coast of Western Australia has long been recognised (Pearce, 1991). By clearly documenting that gradual changes to ocean temperature along the coast are evident at time-scales of days, months and years, this study corroborates and expands the findings of previous studies (Pearce and Feng, 2007; Smale and Wernberg, 2009). Even if complexity is evident, the range of temperature variation between evenly spaced regions supports the contention that the south-western coastline is a suitable natural laboratory, where observations and experiments may be particularly relevant in relation to understanding what ecological changes are likely under projected global warming scenarios (Smale and Wernberg, 2009; Wernberg et al., 2010).

4.2. Habitat structure

The total cover of kelp did not change, but the size of monospecific kelp patches increased. Similarly, total cover of fucooids did not change, but the abundance of mixed kelp–fucooid canopies decreased, and so did the cover of *S. doryocarpa* and ‘other fucooids’, whilst *Sargassum* spp. increased in cover. This indicates a shift in patch structure *per se*, with a homogenisation of kelp patches towards lower latitudes and a switch in

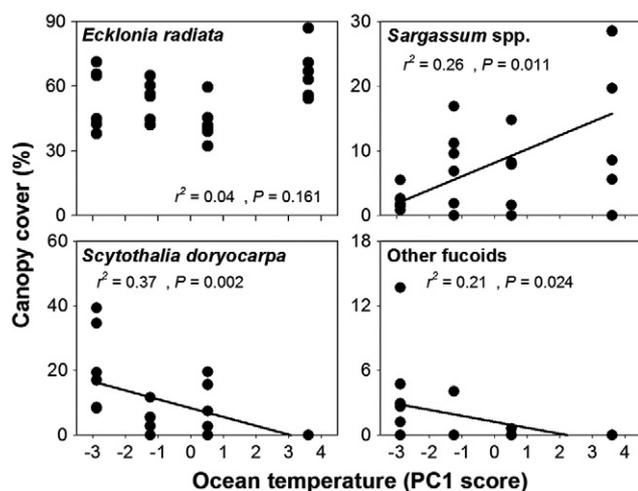


Fig. 4. Relationship between ocean temperature and abundance of dominant canopy-forming taxa. Note, *E. radiata* also includes kelps found in mixed canopies. Ocean temperature incorporates temperature variation at multiple temporal scales (cf. Table 1), and increments correspond to (from left to right) the studied regions (Hamelin Bay, Marmion, Jurien Bay and Kalbarri). Solid lines indicate significant relationships ($P < 0.05$).

species composition of fucoids, which occupy the canopy differently. *S. doryocarpa* is usually found as single or a few individuals interspersed between the kelps, whereas *Sargassum* spp. generally occupies distinct patches, punctuated within the kelp canopy (*personal observation*, T. Wernberg). Several mechanisms could combine to produce this pattern.

The ecological processes that underpin kelp canopy maintenance change along the latitudinal temperature gradient, even if the abundance of *E. radiata* remains the same (Wernberg et al., 2010). In the cooler regions (Hamelin Bay and Marmion), adult kelps suppress the establishment of kelp recruits, leaving more opportunity for tough fucoids such as *S. doryocarpa* to establish and form mixed canopies. In contrast, in the warmer regions (Jurien Bay and Kalbarri), adult kelps reinforce existing kelp patches by facilitating kelp recruitment and recruit growth (Wernberg et al., 2010). This shift from competitive to facilitative adult-recruit interactions could thus promote mixed canopies in cooler water and monospecific kelp canopies in warmer water.

The genus *Scytothalia* is endemic to the south and southwest coast of Australia, and only has the one species (Huisman, 2000). It likely evolved from the temperate Tethyan marine flora on the south coast (Phillips, 2001), and its declining abundance into warmer waters (this study, Smale et al., 2010) is presumably due to its low thermo-tolerance. However, little physiological and ecological work has been done on *S. doryocarpa*, and this hypothesis remains to be tested. *Sargassum* spp., on the other hand, is a very broadly distributed genus with many species of warm-temperate to tropical affinity in Australia (Huisman, 2000). *Sargassum* spp. are competitively inferior to kelps, and in cooler water *Sargassum* spp. patches often revert to kelp within 1–2 years (Toohey et al., 2007). Potentially, the increased predominance of *Sargassum* spp. to the north is promoted by temperature-related suppression of kelp canopy recovery following natural disturbances (Wernberg et al., 2010).

The proportion of canopy-free gaps did not vary systematically along the climate gradient. This supports our initial assumption that there was little difference in the overall disturbance regime amongst regions across the latitudinal gradient, as a previous study found a significant positive relationship between wave exposure and the prevalence of gaps in the canopy cover (Wernberg and Connell, 2008).

In summary, temperature-mediated shifts in ecological interactions within kelp canopies and the identity of dominant fucoids, promote monospecific kelp patches at the expense of mixed canopies. It is important to stress that the increase in monospecific kelp canopies should not be taken as a sign that kelp canopies are impervious to

temperature effects, as recent experiments have shown that the underlying concurrent physiological changes (e.g., Stæhr and Wernberg, 2009), make the canopy less resilient to additional perturbation (Wernberg et al., 2010). Also, mixed canopies are currently a defining feature across vast expanses of the Australian south coast (Connell and Irving, 2008; Wernberg and Connell, 2008; Smale et al., 2010). Because of the southward facing coast line, even a slight southward shift in isotherms could, therefore, result in habitat changes across thousands of kilometres of coastline.

4.3. Important temperature factors

Temperature variables representing both acute, seasonal and inter-annual temperature regimes were found amongst the five best correlates of habitat structure, and all (except temperature at the time of sampling) concerned extreme or adverse conditions. This suggests a complex influence of ocean temperature, driven by differences in the outer temperature envelope rather than mean conditions.

Temperature thresholds can be a major determinant of broad-scale species distributions (Schils and Wilson, 2006). The two temperature variables best correlated to habitat structure were minimum temperature and hot days > 20 °C in the preceding year. Presumably the importance of these variables relates to the recruitment and growth of the habitat formers present at the time of sampling. For temperate algae, reproduction and recruitment are often restricted to substantially lower temperatures than the upper limit for adult survival (Novaczek, 1984; Matson and Edwards, 2007), and growth and productivity have been shown to be adversely affected at high temperatures (Kirkman, 1984; Hatcher et al., 1987). The high correlation with these variables could then be due to a required window where temperatures are low enough to successfully complete reproduction and recruitment, and subsequently enough cool days to attain biomass and canopy cover.

4.4. Non-temperature sources of variation

A prominent feature for all habitat characteristics was the great amount of variation from reef to reef within each region, which in many cases was of equal magnitude to variation across the entire climate gradient. Importantly, this variation does not eliminate the effects of regional temperature regimes, but rather points to a strong influence of additional non-temperature drivers of habitat structure. A range of physical and biological processes could be contributing to this variation.

A study from South Australia, where the same canopy types dominate subtidal reefs, found substantial differences in habitat cover and patch sizes between the seaward and shoreward sides of small islands (Wernberg and Connell, 2008). These patterns were attributed to differences in wave exposure, where high wave exposure was associated with larger patches of all habitat types and kelp dominance, whereas low wave exposure was associated with smaller patches and fucoid dominance. Mixed canopies differed little with wave exposure. With the exception of studies on wave exposure and gap properties, there have been very few direct assessments of the potential drivers of habitat structure on algal dominated reefs. There is abundant evidence to document how other environmental (e.g., reef geology and geomorphology, Harman et al., 2003; Toohey et al., 2007) and ecological (e.g., herbivores and recruitment, Connell and Irving, 2008; Ling, 2008; Wernberg and Connell, 2008) properties can affect the distribution and abundance of habitat-forming organisms. It seems likely that similar environmental and ecological properties have influenced reef to reef differences in habitat structure in this study.

Notwithstanding variability amongst reefs within each region it is important to recognise that the random selection of reef sites within regions was constrained to minimise the influence of non-temperature

drivers between regions (e.g., approximately the same range of depths and wave exposures were targeted). Consequently, to the extent possible, non-temperature drivers of canopy structure did not confound the analysis of the climatic signal.

4.5. Ecological implications

The ecological function of different algal canopies can vary considerably. Due to variation in phenology and thallus structure, kelps and fucoids often present different habitat qualities (e.g., temporal stability, interstitial space, protection from predation, and nutritional value: Wernberg et al., 2001, 2004; Tuya et al., 2008). Monospecific *Ecklonia* canopies, mixed *Ecklonia*–fucalcan canopies and fucalcan canopies have also been shown to modify their immediate environments differently (e.g., light, abrasion, and sedimentation: Jenkins et al., 1999a,b,c; Wernberg et al., 2005; Irving and Connell, 2006b). As both provision of habitat and ecosystem engineering are key mechanisms by which algal canopies affect associated biota, a number of studies have documented how patches of kelp, fucalcan or mixed canopies support different phytal (Wernberg et al., 2004; Tuya et al., 2008), benthic (Goodsell et al., 2004; Irving et al., 2004) and fish (Harman et al., 2003; Tuya et al., 2009) communities.

Metre-scale variation in biogenic habitat structure underpins the uniquely high biodiversity of the Australian west and south coasts (Wernberg et al., 2003; Goodsell et al., 2004; Irving et al., 2004; Smale et al., 2010). Specifically, macroalgal assemblages under monospecific kelp canopies are species poor (Kendrick et al., 1999; Wernberg et al., 2005; Wernberg and Goldberg, 2008) and a loss of macroalgal diversity may follow a reduction in canopy heterogeneity (e.g., Toohey et al., 2007). It is therefore possible, even if overall landscape composition may change little in terms of the dominant habitat-forming species, that an increasing dominance of monospecific kelp canopies, at the expense of mixed canopies, following a shift in dominant fucoids from *Scytothalia* to *Sargassum*, could have significant ecological implications.

5. Conclusion

Significant changes in biogenic habitat structure (increase in monospecific kelp patches, decrease in mixed canopies, and change in fucoid species) were found along a latitudinal gradient in ocean temperature, equivalent to projected temperature increases for the coming 25–50 years. Consequently, the results of this study suggest that, if algal canopies respond to global warming in a similar way as they respond to increasingly warm climate along the studied latitudinal gradient, then the magnitude of projected climate change is great enough to cause a significant reduction in diversity of biogenic habitats on temperate reefs. This is likely to be ecologically significant even if habitat shifts are not catastrophic.

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References

Anderson, M.J., Gorley, R.N., Clarke, K.R., 2008. PERMANOVA + for PRIMER: Guide to Software and Statistical Methods, Vol. PRIMER-E Ltd, Plymouth, UK.
 Blight, A.J., Thompson, R.C., 2008. Epibiont species richness varies between holdfasts of a northern and a southerly distributed kelp species. *J. Mar. Biol. Assoc. UK* 88, 469–475.
 Clarke, K.R., Gorley, R.N., 2006. Primer v6: User Manual/Tutorial, Vol. PRIMER-E Ltd, Plymouth, U.K.
 Coleman, M.A., Gillanders, B.M., Connell, S.D., 2009. Dispersal and gene flow in the habitat-forming kelp, *Ecklonia radiata*: relative degrees of isolation across an east-west coastline. *Mar. Freshw. Res.* 60, 802–809.

Connell, S.D., Irving, A.D., 2008. Integrating ecology with biogeography using landscape characteristics: a case study of subtidal habitat across continental Australia. *J. Biogeogr.* 35, 1608–1621.
 Dayton, P.K., 1972. Towards an understanding of community resilience and the potential effects of enrichment to the benthos at McMurdo Sound, Antarctica. In: Parker, B.C. (Ed.), *Colloquium on Conservation Problems in Antarctica*. Allen Press, Lawrence, Kansas, USA, pp. 81–95.
 Dayton, P.K., Tegner, M.J., Edwards, P.B., Riser, K.L., 1999. Temporal and spatial scales of kelp demography: the role of oceanographic climate. *Ecol. Monogr.* 69, 219–250.
 Edwards, M.S., Estes, J.A., 2006. Catastrophe, recovery and range limitation in NE Pacific kelp forests: a large-scale perspective. *Mar. Ecol. Prog. Ser.* 320, 79–87.
 Goodsell, P.J., Fowler-Walker, M.J., Gillanders, B.M., Connell, S.D., 2004. Variations in the configuration of algae in subtidal forests: implications for invertebrate assemblages. *Aust. Ecol.* 29, 350–357.
 Graham, M.H., 2004. Effects of local deforestation on the diversity and structure of southern California giant kelp forest food webs. *Ecosystems* 7, 341–357.
 Harley, C.D.G., Randall Hughes, A., Hultgren, K.M., Miner, B.G., Sorte, C.J.B., Thornber, C.S., Rodriguez, L.F., Tomanek, L., Williams, S.L., 2006. The impacts of climate change in coastal marine systems. *Ecol. Lett.* 9, 228–241.
 Harman, N., Harvey, E.S., Kendrick, G.A., 2003. Differences in fish assemblages from different reef habitats at Hamelin Bay, south-western Australia. *Mar. Freshw. Res.* 54, 177–184.
 Hatcher, B.G., Kirkman, H., Wood, W.F., 1987. Growth of the kelp *Ecklonia radiata* near the northern limit of its range in Western Australia. *Mar. Biol.* 95, 63–72.
 Hawkins, S.J., Southward, A.J., Genner, M.J., 2003. Detection of environmental change in a marine ecosystem—evidence from the western English Channel. *Sci. Total Environ.* 310, 245–256.
 Hawkins, S.J., Moore, P., Burrows, M.T., Polczanska, E., Mieszowska, N., Jenkins, S.R., Thompson, R.C., Genner, M.J., Southward, A.J., 2008. Complex interactions in a rapidly changing world: responses of rocky shore communities to recent climate change. *Clim. Res.* 37, 123–133.
 Hawkins, S.J., Sugden, H.E., Mieszowska, N., Moore, P., Polczanska, E., Leaper, R., Herbert, R.J.H., Genner, M.J., Moschella, P.S., Thompson, R.C., Jenkins, S.R., Southward, A.J., Burrows, M.T., 2009. Consequences of climate driven biodiversity changes for ecosystem functioning of North European Rocky Shores. *Mar. Ecol. Prog. Ser.* 396, 245–259.
 Huismans, J.M., 2000. Marine Plants of Australia. Vol. University of Western Australia Press, Perth, p. 300.
 Irving, A.D., Connell, S.D., 2006a. Physical disturbance by kelp abrades erect algae from the understorey. *Mar. Ecol. Prog. Ser.* 324, 127–137.
 Irving, A.D., Connell, S.D., 2006b. Predicting understorey structure from the presence and composition of canopies: an assembly rule for marine algae. *Oecologia* 148, 491–502.
 Irving, A.D., Connell, S.D., Gillanders, B.M., 2004. Local complexity in patterns of canopy-benthos associations produces regional patterns across temperate Australasia. *Mar. Biol.* 144, 361–368.
 Jenkins, S.R., Norton, T.A., 1999a. Interaction between a fucoid canopy and limpet grazing in structuring a low shore intertidal community. *J. Exp. Mar. Biol. Ecol.* 233, 41–63.
 Jenkins, S.R., Norton, T.A., Hawkins, S.J., 1999b. Settlement and post-settlement interactions between *Semibalanus balanoides* (L.) (Crustacea: Cirripedia) and three species of fucoid canopy algae. *J. Exp. Mar. Biol. Ecol.* 236, 49–67.
 Jenkins, S.R., Norton, T.A., Hawkins, S.J., 1999c. Interactions between canopy forming algae in the eulittoral zone of sheltered rocky shores on the Isle of Man. *J. Mar. Biol. Assoc. UK* 79, 341–349.
 Kendall, M.A., Burrows, M.T., Southward, A.J., Hawkins, S.J., 2004. Predicting the effects of marine climate change on the invertebrate prey of the birds of rocky shores. *Ibis* 146, 40–47.
 Kendrick, G.A., Walker, D.I., 1991. Dispersal distances for propagules of *Sargassum spinuligerum* (Sargassaceae Phaeophyta) measured directly by vital staining and venturi suction sampling. *Mar. Ecol. Prog. Ser.* 79, 133–138.
 Kendrick, G.A., Lavery, P.S., Phillips, J.C., 1999. Influence of *Ecklonia radiata* kelp canopy structure on macro-algal assemblages in Marmion Lagoon, Western Australia. *Hydrobiologia* 399, 275–283.
 Kennelly, S.J., 1987. Physical disturbances in an Australian kelp community. I. Temporal effects. *Mar. Ecol. Prog. Ser.* 40, 145–153.
 Kirkman, H., 1984. Standing stock and production of *Ecklonia radiata* (C.Ag.) J. Agardh. *J. Exp. Mar. Biol. Ecol.* 76, 119–130.
 Lima, F.P., Ribeiro, P.A., Queiroz, N., Hawkins, S.J., Santos, A.M., 2007. Do distributional shifts of northern and southern species of algae match the warming pattern? *Glob. Change Biol.* 13, 2592–2604.
 Ling, S., 2008. Range expansion of a habitat-modifying species leads to loss of taxonomic diversity: a new and impoverished reef state. *Oecologia* 156, 883–894.
 Ling, S.D., Johnson, C.R., Frusher, S.D., Ridgway, K.R., 2010. Overfishing reduces resilience of kelp beds to climate-driven catastrophic phase shift. *Proc. Natl Acad. Sci. USA* 106, 22341–22345.
 Lourey, M.J., Dunn, J.R., Waring, J., 2006. A mixed-layer nutrient climatology of Leeuwin Current and Western Australian shelf waters: seasonal nutrient dynamics and biomass. *J. Mar. Syst.* 59, 25–51.
 Martinez, E.A., Cardenas, L., Pinto, R., 2003. Recovery and genetic diversity of the intertidal kelp *Lessonia nigrescens* (Phaeophyceae) 20 years after *El Niño* 1982/83. *J. Phycol.* 39, 504–508.
 Matson, P., Edwards, M., 2007. Effects of ocean temperature on the southern range limits of two understorey kelps, *Pterygophora californica* and *Eisenia arborea*, at multiple life-stages. *Mar. Biol.* 151, 1941–1949.
 Merzouk, A., Johnson, L.E., 2011. Kelp distribution in the northwest Atlantic Ocean under a changing climate. *J. Exp. Mar. Biol. Ecol.* 400 xx–xx.

- Moore, P., Hawkins, S.J., Thompson, R.C., 2007. Role of biological habitat amelioration in altering the relative responses of congeneric species to climate change. *Mar. Ecol. Prog. Ser.* 334, 11–19.
- Novaczek, I., 1984. Response of gametophytes of *Ecklonia radiata* (Laminariales) to temperature in saturating light. *Mar. Biol.* 82, 241–245.
- Parmesan, C., Yohe, G., 2003. A globally coherent fingerprint of climate change impacts across natural systems. *Nature* 421, 37–42.
- Pearce, A.F., 1991. Eastern boundary currents of the southern hemisphere. *J. R. Soc. W Aust.* 74, 35–45.
- Pearce, A., Feng, M., 2007. Observations of warming on the Western Australian continental shelf. *Mar. Freshw. Res.* 58, 914–920.
- Phillips, J., 2001. Marine macroalgal biodiversity hotspots: why is there high species richness and endemism in southern Australian marine benthic flora? *Biodivers. Conserv.* 10, 1555–1577.
- Pimm, S.L., 2008. Biodiversity: climate change or habitat loss – which will kill more species? *Curr. Biol.* 18, 117–119.
- Poloczanska, E.S., Babcock, R.C., Butler, A., Hobday, A.J., Hoegh-Guldberg, O., Kunz, T.J., Matear, R., Milton, D.A., Okey, T.A., Richardson, A.J., 2007. Climate change and Australian marine life. *Oceanogr. Mar. Biol. Ann. Rev.* 45, 407–478.
- Ridgway, K.R., 2007. Long-term trend and decadal variability of the southward penetration of the East Australian Current. *Geophys. Res. Lett.* 34.
- Rosenzweig, C., Karoly, D.J., Vicarelli, M., Neofotis, P., Wu, Q., Casassa, G., Menzel, A., Root, T.L., Estrella, N., Seguin, B., Tryjanowski, P., Liu, C., Rawlins, S., Imeson, A., 2008. Attributing physical and biological impacts to anthropogenic climate change. *Nature* 453, 353–357.
- Schiel, D.R., 2011. Biogeographic patterns and long-term changes on New Zealand coastal reefs: non-trophic cascades from diffuse and local impacts. *J. Exp. Mar. Biol. Ecol.* 400, 33–51 (this issue).
- Schiel, D.R., Foster, M.S., 1986. The structure of subtidal algal stands in temperate waters. *Oceanogr. Mar. Biol. Ann. Rev.* 24, 265–307.
- Schiel, D.R., Steinbeck, J.R., Foster, M.S., 2004. Ten years of induced ocean warming causes comprehensive changes in marine benthic communities. *Ecology* 85, 1833–1839.
- Schils, T., Wilson, S.C., 2006. Temperature threshold as a biogeographic barrier in northern Indian Ocean macroalgae. *J. Phycol.* 42, 749–756.
- Smale, D.A., Wernberg, T., 2009. Satellite-derived SST data as a proxy for water temperature in near-shore benthic ecology. *Mar. Ecol. Prog. Ser.* 387, 27–37.
- Smale, D.A., Kendrick, G.A., Wernberg, T., 2010. Assemblage turnover and taxonomic sufficiency of subtidal macroalgae at multiple spatial scales. *J. Exp. Mar. Biol. Ecol.* 384, 76–86.
- Southward, A.J., Hawkins, S.J., Burrows, M.T., 1995. Seventy years' observations of changes in distribution and abundance of zooplankton and intertidal organisms in the western English Channel in relation to rising sea temperature. *J. Therm. Biol.* 20 (1–2), 127–155.
- Stæhr, P.A., Wernberg, T., 2009. Physiological responses of *Ecklonia radiata* (Laminariales) to a latitudinal gradient in ocean temperature. *J. Phycol.* 45, 91–99.
- Thompson, R.C., Wilson, B.J., Tobin, M.L., Hill, A.S., Hawkins, S.J., 1996. Biologically generated habitat provision and diversity of rocky shore organisms at a hierarchy of spatial scales. *J. Exp. Mar. Biol. Ecol.* 202, 73–84.
- Thomsen, M.S., Wernberg, T., Altieri, A., Tuya, F., Gulbransen, D., McGlathery, K., Holmer, M., Silliman, B.R., 2010. Habitat cascades: the conceptual context and global relevance of facilitation cascades via habitat formation and modification. *Integr. Comp. Biol.* 50, 158–175.
- Toohy, B.D., Kendrick, G.A., Harvey, E.S., 2007. Disturbance and reef topography maintain high local diversity in *Ecklonia radiata* kelp forests. *Oikos* 116, 1618–1630.
- Tuya, F., Wernberg, T., Thomsen, M.S., 2008. The spatial arrangement of reefs alters the ecological patterns of fauna between interspersed algal habitats. *Estuar. Coast. Shelf Sci.* 78, 774–782.
- Tuya, F., Wernberg, T., Thomsen, M., 2009. Habitat structure affect abundances of labrid fishes across temperate reefs in south-western Australia. *Environ. Biol. Fish.* 86, 311–319.
- Vanderklift, M.A., Kendrick, G.A., 2005. Contrasting influence of sea urchins on attached and drift macroalgae. *Mar. Ecol. Prog. Ser.* 299, 101–110.
- Vanderklift, M.A., Lavery, P.S., Waddington, K.I., 2009. Intensity of herbivory on kelp by fish and sea urchins differs between inshore and offshore reefs. *Mar. Ecol. Prog. Ser.* 376, 203–211.
- Walther, G.-R., Post, E., Coney, P., Menzel, A., Parmesan, C., Beebee, T.J.C., Fromentin, J.-M., Hoegh-Guldberg, O., Bairlein, F., 2002. Ecological responses to recent climate change. *Nature* 416, 389–395.
- Wernberg, T., 2005. Holdfast aggregation in relation to morphology, age, attachment and drag for the kelp *Ecklonia radiata*. *Aquat. Bot.* 82, 168–180.
- Wernberg, T., 2006. Scale of impact determines early post-disturbance assemblage structure in subtidal *Fucus* beds in the Baltic Sea (Bornholm, Denmark). *Eur. J. Phycol.* 41, 105–113.
- Wernberg, T., Connell, S.D., 2008. Physical disturbance and subtidal habitat structure on open rocky coasts: effects of wave exposure, extent and intensity. *J. Sea Res.* 59, 237–248.
- Wernberg, T., Goldberg, N., 2008. Short-term temporal dynamics of algal species in a subtidal kelp bed in relation to changes in environmental conditions and canopy biomass. *Estuar. Coast. Shelf Sci.* 76, 265–272.
- Wernberg, T., Thomsen, M.S., Stæhr, P.A., Pedersen, M.F., 2001. Comparative phenology of *Sargassum muticum* and *Halidrys siliquosa* (Phaeophyceae : Fucales) in Limfjorden, Denmark. *Bot. Mar.* 44 (1), 31–39.
- Wernberg, T., Kendrick, G.A., Phillips, J.C., 2003. Regional differences in kelp-associated algal assemblages on temperate limestone reefs in south-western Australia. *Divers. Distrib.* 9, 427–441.
- Wernberg, T., Thomsen, M.S., Stæhr, P.A., Pedersen, M.F., 2004. Epibiota communities of the introduced and indigenous macroalgal relatives *Sargassum muticum* and *Halidrys siliquosa* in Limfjorden (Denmark), Helgoland. *Mar. Res.* 58, 154–161.
- Wernberg, T., Kendrick, G.A., Toohey, B.D., 2005. Modification of the physical environment by an *Ecklonia radiata* (Laminariales) canopy and implications for associated foliose algae. *Aquat. Ecol.* 39, 419–430.
- Wernberg, T., White, M., Vanderklift, M.A., 2008. Population structure of turbinid gastropods on wave-exposed subtidal reefs: effects of density, body size and algae on grazing behaviour. *Mar. Ecol. Prog. Ser.* 362, 169–179.
- Wernberg, T., Thomsen, M.S., Tuya, F., Kendrick, G.A., Stæhr, P.A., Toohey, B.D., 2010. Decreasing resilience of kelp beds along a latitudinal temperature gradient: potential implications for a warmer future. *Ecol. Lett.* 13, 685–694.
- Wernberg, T., Russell, B.D., Moore, P.J., Ling, S.D., Smale, D.A., Campbell, A., Coleman, M.A., Steinberg, P.D., Kendrick, G.A., Connell, S.D., 2011. Impacts of climate change in a global hotspot for temperate marine biodiversity and ocean warming. *J. Exp. Mar. Biol. Ecol.* 400 xx–xx.
- Zar, J.H., 1996. *Biostatistical Analysis*. Prentice Hall, Upper Saddle River, p. 662.